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1 An Approach to Improve Direct Runoff Estimates and Reduce Uncertainty in

2 the Calculated Groundwater Component in Water Balances of Large Lakes

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21 Abstract

22 Groundwater is important in the overall water budget of a lake because it affects the 23 quantity and quality of surface water and the ecological health of the lake. The water balance 24 equation is frequently used to estimate the net groundwater flow for small lakes but is seldom 25 used to determine net groundwater flow components for large lakes because: 1) errors 26 accumulate in the calculated groundwater term, and 2) there is an inability to accurately quantify 27 the direct runoff component. In this water balance study of Lake Pyhäjärvi (155 km²) in Finland, 28 it was hypothesized a hydrograph separation model could be used to estimate direct runoff to the 29 lake and, when combined with a rigorous uncertainty analyses, would provide reliable net 30 groundwater flow estimates. The PART hydrograph separation model was used to estimate 31 annual per unit area direct runoff values for the watershed of the inflowing Yläneenjoki River (a 32 subwatershed of the lake) which were then applied to other physically similar subwatersheds of 33 the lake to estimate total direct runoff to the lake. The hydrograph separation method provided 34 superior results and had lower uncertainty than the common approach of using a runoff 35 coefficient based method. The average net groundwater flow into the lake was calculated to be 36 +43 mm per year (+3.0% of average total inflow) for the 38 water years 1971 to 2008. It varied 37 from -197 mm to 284 mm over that time, and had a magnitude greater than the uncertainty for 17 38 of the 38 years. The average indirect groundwater contribution to the lake (i.e., the groundwater 39 part of the inflowing rivers) was 454 mm per year (+32% of average total inflow) and 40 demonstrates the overall importance of groundwater. The techniques in this study are applicable 41 to other large lakes and may allow small net groundwater flows to be reliably quantified in 42 settings that might otherwise be unquantifiable or completely lost in large uncertainties.

43 Keywords: groundwater, direct runoff, lake, water budget, uncertainty

44 **1. Introduction**

45 The flow of groundwater into lakes is important because it can affect: the quantity and 46 quality of the surface water (LaBaugh et al., 1995; Winter, 1999; Dubrovsky et al., 2010; Fruh, 47 1967; Bruce et al., 2009); the ecosystem health (Hayashi and Rosenberry, 2002); the distribution 48 of aquatic life (Baird and Wilby, 1999; Rosenberry et al., 2000); and the quality of the fish 49 habitat (Power et al., 1999). Estimates of net groundwater discharge to a lake can indicate the 50 relative importance of groundwater in the water budget, but accurately quantifying total 51 discharge can be a challenge. Groundwater flows into and out of lakes can be estimated using: 52 direct point measurements of flow (Cartwright et al., 1979; Cherkauer and Nader, 1989; Harvey 53 et al., 1997 and 2000); water balance calculations (Winter, 1981; Sacks et al., 1998; Zacharias et 54 al., 2003); isotopic tracers (Walker and Krabbenhoft, 1998; Stets et al., 2010), and numerical 55 modeling of the lake and its watershed (Feinstein et al., 2010; Hoaglund et al., 2002; Mylopoulos 56 et al., 2007). Point measurement techniques are useful but impractical to employ on a lake-wide 57 basis, particularly when the lake is large and there is substantial spatial heterogeneity in lakebed 58 deposits and flows. Likewise, geochemical methods are difficult to use in large lakes because of 59 spatial variability in water quality and challenges in defining appropriate end member 60 concentrations for calculating mixing ratios. Numerical models that quantify groundwater flow 61 are potentially very useful and can handle considerable spatial and temporal complexities; 62 however, the lack of field data to constrain and populate these models generally results in major 63 simplifying assumptions which produce uncertainties and errors that are either unknown or not 64 readily quantifiable. The water balance method requires the quantification of inflows (precipitation, direct runoff, surface water inflows), outflows (evaporation, surface water 65 outflows), and change in lake storage to calculate net groundwater flow. If properly done, the 66

water balance equation has the potential to provide accurate estimates of the net groundwater
flow (i.e., groundwater inflow minus groundwater outflow, which represents a minimum value
for groundwater discharge) with potentially less effort and uncertainty than is associated with the
other techniques. Despite this potential, the water balance method tends not to be used to
determine net groundwater discharges for large lakes (Quinn and Guerra, 1986; Neff and Killian,
2003; Lenters, 2004; Neff and Nicholas, 2005).

73 There are two main reasons why water balances performed on large lakes do not attempt 74 to quantify groundwater-surface water exchanges and, instead, either assume groundwater 75 contributions are insignificant (i.e., are zero) or simply lump them together with the direct runoff 76 into a combined runoff term. The first reason is that net groundwater flow is usually solved for as 77 an unknown in the water balance equation, which means all the uncertainty in other components 78 translates to and accumulates in the uncertainty of the groundwater component. Even what 79 appear to be small relative errors on large components (e.g., precipitation or evaporation) may 80 result in errors of substantial absolute magnitude that are larger than the groundwater component 81 being quantified (Winter, 1981; Thodal, 1997). Unfortunately, many studies do not perform the 82 uncertainty analysis necessary to assess the reliability of results even though several studies 83 discuss how to quantify uncertainties (Winter, 1981; Lee and Swancar, 1997; Winter and 84 Rosenberry, 2009; Neff and Nicholas, 2005). Even in studies where the net groundwater flow in 85 the water budget as a percent of total inflow appeared to be important (e.g., Zacharias et al., 86 2003; Demlie et al., 2007; and Avenew and Gebreegziabher, 2006), uncertainty analysis of the 87 groundwater term has not been included. Without the uncertainty analyses, it is not known if the 88 calculated values of net groundwater flow are accurate and representative.

89 The second reason why net groundwater discharge is not calculated for lakes is because it 90 requires the direct runoff component (i.e., non-channelized overland flow and interflow) be 91 quantified and this is often neglected or cannot be done with confidence or certainty due to a lack 92 of suitable methods. The direct runoff component is usually ignored for large lakes (Neff and 93 Nicholas, 2005; Lenters, 2004; Neff and Killian, 2003), and little work has been done in the last 94 three decades to specifically estimate non-channelized runoff to lakes despite its inclusion in 95 data-intensive time-stepping models such as SWAT (e.g., Menking et al., 2003), MOD-HMS 96 (e.g., Panday and Huyakorn, 2004), and WATLAC (e.g., Zhang, 2011). The few methods that 97 have been applied have been for small lakes and were originally developed for streams. The 98 methods include: the curve number (CN) method (Natural Resources Conservation Service, 99 2004; Motz et al., 2001), the use of coefficients associated with varying land use and 100 permeability (Sacks et al., 1998; Dames and Moore, 1992), and the extrapolation of hydrograph 101 separation results (Newbury and Beaty, 1980; Schindler et al., 1976). The hydrograph separation 102 model approach is appealing because it represents an empirical relationship derived from and 103 calibrated to a portion of that particular lake's watershed and takes into account the actual 104 physical and climatological conditions at the site without relying on models that extrapolate and 105 use empirical runoff relationships derived at other sites with different conditions. The 106 hydrograph separation method has not been applied to large lakes, and there is a need to 107 determine its applicability and accuracy when applied to large lakes.

108 An opportunity to examine these issues concerning quantification of net groundwater 109 discharge and direct runoff to large lakes was presented when concerns were expressed regarding 110 the current and future water quality of Lake Pyhäjärvi (155 km²), located in glacial terrain near 111 Säkylä, Finland. The concerns focused on the eutrophication of the lake resulting in part from the

112 effects of the agricultural watershed around the lake, along with impacts on the fishing industry, 113 recreational enjoyment, and overall ecological integrity of the lake (Kirkkala, 2014). Early 114 studies of the lake (Hyvärinen et al., 1973; Kuusisto, 1975; Järvinen, 1978; Eronen et al., 1982) 115 either insufficiently assessed the net groundwater component of the lake's water budget or 116 assumed it was negligible (i.e., zero); however, recent work indicated significant groundwater 117 discharge might occur through an esker that intersects Lake Pyhäjärvi and at other specific 118 locations along the shoreline (Rautio, 2009; Korkka-Niemi et al., 2011; Rautio and Korkka-119 Niemi, 2011). Moreover, indirect groundwater discharge, where groundwater discharges to a 120 river and then is transported into the lake by the river, can also influence the quantity and quality 121 of water in large lakes (Holtschlag and Nicholas, 1998; Neff et al., 2005). It was hypothesized 122 that using historical climatological and hydrological data, a carefully conducted water balance 123 study could be used to successfully estimate the net groundwater flow into the lake, provided that 124 a rigorous uncertainty analysis was performed to characterize potential errors and that a suitable 125 method for determining direct runoff could be used. A specific objective of this study was to 126 evaluate whether a hydrograph separation method that has been applied to streams and small 127 lakes to estimate direct runoff could be successfully applied to a large lake. This study 1) 128 provides the first rigorous water balance and estimates of net groundwater flow and indirect 129 groundwater discharge for Lake Pyhäjärvi, 2) demonstrates the importance of uncertainty 130 analyses, and 3) successfully tests the hypothesis that using a hydrograph separation method to 131 estimate the direct runoff component to a large lake is a viable approach for water balances. This 132 approach could be applicable to other large water bodies in various landscape settings.

133

134 **2. Background**

135 Lake Pyhäjärvi (60°54′-61°06′N, 22°09′-22°25′E) is the largest lake in southwestern 136 Finland (155 km²) and is a valuable fishery and recreational area (Ventelä et al., 2007; Ventelä et 137 al., 2005). The lake is guite shallow (5.4 m on average) with a maximum depth of 26 m (e.g., 138 Kirkkala, 2014), and it makes up a large percentage (25%) of its watershed (Figure 1). Lake Pyhäjärvi's watershed (616 km²) is predominately agricultural land (Luoto, 2000; Häkkinen, 139 140 1996). The ground elevations in the watershed range from about 40 to 145 masl, and it is 141 relatively flat with an average topographic slope of 2.8% (MML, 2009c; ESRI, 2010). Two 142 rivers (Yläneenjoki and Pyhäjoki) are gauged, drain the agricultural lands in the south and east, 143 and flow into the lake; while one river (Eurajoki, also gauged) flows from the northern end of the lake at Kauttua Falls and flows to the Baltic Sea. The remaining area (304 km²) of the lake's 144 145 watershed is ungauged and consists of four subwatersheds with single channels that drain water 146 into the lake and another six subwatersheds that do not have significant drains or channels 147 (Figure 1).

148 The landscape around Lake Pyhäjärvi has been sculpted by glacial erosion and 149 deposition. The surficial geology around the lake is shown in Figure 2 and consists primarily of 150 thin, discontinuous till layers, numerous granite and sandstone bedrock outcrops, and to a lesser 151 extent clays, peats, and silts. Figure 3 shows that the watershed contains very few coarse grained 152 aquifer deposits. Among these is the Kuivalahti-Säkylä esker, which is connected to the large 153 Säkylänharju-Virttaankangas Glaciofluvial Complex that lies mostly outside the watershed and is 154 on the eastern side of the Pyhäjoki River's subwatershed. The esker is found along 15 km of the 155 lake's northeastern shoreline and contains several aquifers, including the Honkala Aquifer (Figure 3). Figure 2 shows that the Yläneenjoki River's subwatershed (234 km²) contains more 156

clay and bedrock and is less permeable than the Pyhäjoki River's subwatershed (78 km²), which
contains sands and coarse-grained materials of the Virttaankangas Glaciofluvial Complex
(Eronen et al., 1982).

160 Little is known about the groundwater-surface water interactions for Lake Pyhäjärvi. 161 Because the Eurajoki River drains water from the lake, the lake may be a gaining lake (i.e., gains 162 groundwater), but for many lakes (referred to as flow-through lakes) groundwater can enter the 163 lake in one area as groundwater discharge and surface water can leave the lake through the 164 bottom sediments as groundwater recharge at another location. Groundwater discharge to the 165 lake has been documented at specific locations along the shoreline (Rautio, 2009; Korkka-Niemi 166 et al., 2011; Rautio and Korkka-Niemi, 2011; Artimo, 2002), but areas of groundwater recharge 167 have not been documented. Hydraulic head data for wells in the watershed (Artimo, unpublished 168 report, 1998; Artimo, 2002; Wiebe, 2012) show that groundwater hydraulic head gradients 169 indicate flow toward the lake, even at the northern end of the lake where the Eurajoki River exits 170 the lake and groundwater recharge conditions might be anticipated. It is not known if 171 groundwater-surface water exchanges occur beneath the lake with the underlying Rapakivi 172 granite and Satakunta sandstone, but the bedrock generally has low permeability and the 173 exchange would likely need to involve regional or intermediate groundwater flow systems. This 174 study was undertaken to provide insight regarding the importance of groundwater with regards to 175 the lake by estimating the net groundwater flow component within the overall water balance of 176 the lake.

177

178 **3. Methods**

179 3.1 Water Balance Approach

A water balance equation may be used either to solve for an unknown component such as evaporation or to verify that estimated input and output components balance (Järvinen, 1978; Sacks et al., 1998; Lenters, 2004; Trask, 2007). For the Lake Pyhäjärvi study, the net groundwater flow was estimated by solving the equation for a defined time period in which all other inputs and outputs, and the changes in storage volume, were known. Operating on a lake area basis over a single water year (October 1 to September 30), the water balance equation for the lake is:

,

(1)

187 where: *G* represents the net groundwater flow into the lake, h_S is the vertical change in lake stage 188 (increases being positive), E is the sum of evaporative losses from the lake, W is the amount 189 withdrawn from the lake by pumping, *P* is the amount of direct precipitation on the lake, *R* is the 190 normalized net river flow (sum of inflowing minus outflowing) plus channelized flow into the 191 lake from ungauged subwatersheds, and *DR* is the normalized direct runoff contribution (non-192 channelized overland flow and interflow) from the watershed into the lake. Normalization of the 193 *R* and *DR* components consisted of dividing their total volume of water for the year by the area 194 of the lake to obtain values in mm per year. Unless otherwise stated in this paper, water balance 195 components that are expressed in mm per year are normalized values equivalent to a volume per 196 unit lake area per water year. Using this equation, the net groundwater flow was estimated for 38 197 water years (October to September) between 1971 and 2009, which allowed the method to be 198 evaluated for a variety of different climatological conditions and to examine temporal trends in

results. Resulting positive values of *G* represent net groundwater discharge conditions (i.e., on
the whole the surface water is gaining groundwater) and negative values of *G* represent net
groundwater recharge (i.e., on the whole the lake is losing surface water to groundwater), but in
each case, there may be both gaining and losing portions of the lake.

203

204 3.2 Quantifying Uncertainties

The solving of the water balance equation for the net groundwater flow means that all the errors associated with each of the other components accumulate in the error associated with the net groundwater term. A standard method for calculating water balance uncertainty (described by Winter, 1981; Lee and Swancar, 1997; and Sacks et al., 1998) was used to determine the uncertainty for each water year:

/ , (2)

210 where δG is the (absolute) uncertainty estimate for the net groundwater flow into the lake, and 211 $\delta h_{S}, \delta E, \delta W, \delta P, \delta R$, and δDR are the absolute uncertainty estimates associated with the lake 212 water level change and other terms in (1), respectively. The absolute uncertainty for each 213 component may also be composed of multiple uncertainties (e.g., equipment measurement errors 214 and data interpolation errors), which in turn are calculated using equations similar in form to 215 Equation 2 (see Tyler, 1977; Ramette, 1981; Taylor, 1997; and Lee and Swancar, 1997). The 216 common assumption (Winter, 1981) that the sources of uncertainty for each component are 217 independent was made. Table 1 lists all the uncertainties that contribute to each individual 218 component and shows the equations and data used to calculate the absolute or relative 219 uncertainty value for each particular component.

221 **3.3 Water Balance Components**

222 **3.3.1. Lake Storage**

223 The change in lake storage (h_s) over the course of each water year was calculated from 224 daily water level measurements (OIVA/HERTTA, 8 Sep 2010) taken at the staff gauge at the 225 north end of the lake. The difference between the lake stage at the start of two consecutive water 226 years constituted the change in storage for the water year thus bracketed (normalized by the lake 227 area by default). The absolute uncertainty (δh_S) for the storage was calculated (Table 1) using an 228 uncertainty of \pm 5.0 mm associated with the staff gauge measurement and an uncertainty value of 229 \pm 25 mm to account for half of the possible lake stage fluctuations that could be caused by wind 230 driven seiche effects (Hyvärinen et al., 1973). The δh_s value was representative of fluctuations 231 observed during the days before the start of each water year.

232 **3.3.2.** Evaporation

233 The lake evaporation (*E*) was estimated using two different types of data, depending on 234 the time of year. For the months of May to November, data from a Class A evaporation pan 235 located at the Jokioinen meteorological station (about 60 km SE of the lake) was used 236 (OIVA/HERTTA, 5 Jun 2010). An average pan coefficient of 0.80 was assigned, which was 237 consistent with the only pan coefficients available in the region (i.e., three years of data for 238 summer months at three meteorological stations within 60 km of the lake); these were between 239 0.76 and 1.25 (Järvinen, 1978). The value is consistent with coefficients from other studies of 240 regions near oceans in the United States (Hounam, 1973; Kohler et al., 1959). The uncertainty in 241 the evaporation estimates for each year was assigned to be 15%, based on the estimated accuracy

242 range associated with a pan coefficient that accounts for this type of lake depth and climatic 243 regime (Dingman, 1994; Harbeck et al., 1954). For the five months when pan data were not 244 available due to freezing conditions (December to April), an evaporation rate of 8 mm per month 245 was assigned based on work on Lake Pyhäjärvi that was performed by Kuusisto (1975), who 246 employed the Dalton-type formula developed by Shuliakovski (1969) to obtain the value. The 247 uncertainty associated with the Dalton type measurements was assumed to be 15% for 248 consistency with the evaporation estimates for the other seven months of the year. The absolute 249 uncertainty for evaporation (δE) was calculated as shown in Table 1.

250

3.3.3. Pumping Withdrawals

251 Few data were available regarding the total amount of water pumped from the lake each 252 year (*W*). The Lohiluoma pumping station, which is located beside the lake near its northern 253 extent, has a municipal well that reportedly extracted (by induced infiltration) 4700 m³/d from 254 the lake during 2010 (J. Reko, pers. comm., 2010). Historical data were not available, although 255 the well has been in operation since 1965 (OIVA/HERTTA, 15 Jun 2011). Lake water extraction 256 for irrigation was not included due to lack of data, nor were other known but minor withdrawals 257 included. To account for the possible variations in annual pumping rate and for the minor withdrawals, an uncertainty of $\pm 500 \text{ m}^3/\text{d}$ was assigned. 258

259 3.3.4. Precipitation

The direct precipitation on the lake (*P*) was estimated for each water year using data from eight nearby meteorological stations that are within 70 km of the lake (Finnish Meteorological Institute [FMI], 24 May 2011; OIVA/HERTTA, 13 Oct 2010). The isohyetal method (e.g., Dingman, 1994) was employed to spatially extrapolate and estimate precipitation over the lake for each water year using the available data. Surfer 8 (Golden Software, Inc., 2002) was used to

265 contour the point precipitation sums for each water year via point kriging for a region 95 km E-266 W by 59 km N-S that encompassed the lake. ArcMap 10.0 GIS software (ESRI, 2010) was used 267 to calculate the areas between 2 mm contour intervals. The absolute uncertainty of the 268 precipitation (δP) was calculated (Table 1) using a baseline value of 5.0% for the instrument 269 error (Winter, 1981) and a year to year spatial interpolation error term. The spatial interpolation 270 term ranged from 0.2 to 15% and was the absolute value of the difference between the magnitude 271 of total precipitation estimated for the lake using the isohyetal method and the magnitude 272 obtained using a second spatial interpolation method (an areal average method using an 273 arithmetic mean value; OIVA/HERTTA, 6 Oct 2010) for the watershed. This areal average 274 approach is outlined in Winter (1981), who cites Linsley et al. (1958).

275 **3.3.5. River Discharge**

276 River discharge estimates were compiled from the net river discharge into the lake from 277 the three gauged rivers, and from per unit area river flow extrapolations from the Yläneenjoki River for the four ungauged subwatersheds with single channel drainage (i.e., $R = R_{net} + R_{single}$ 278 279 _{chan}). River discharge estimates were obtained for the Yläneenjoki, Pyhäjoki, and Eurajoki Rivers 280 using gauging station flow estimates based on rating curves for daily water level measurements 281 at weirs (OIVA/HERTTA, 23 Sep 2010). The sum of the two inflowing rivers minus the 282 outflowing river yielded the net river discharge (R_{net}). Because the gauges for the two inflowing rivers were located a short distance upstream of the confluences of the rivers with the lake 283 284 (Figure 1), the total flows for the rivers were corrected (adjusted upwards) to account for 285 contributions from the ungauged part of the river's watershed. In order to do this, river flow per 286 unit gauged area was multiplied by the area of the ungauged portion and added to the flow for 287 the river prior to calculating R_{net} (described by Wiebe, 2012). Similarly, Yläneenjoki flow per

unit gauged area was applied to the areas of the four single drainage channel subwatersheds 288 289 (*R*_{single chan}). Sums (*R*) for each water year were normalized by dividing values by the average 290 lake area. Groundwater discharge into the two inflowing rivers was included in the flow 291 volumes, and also in the per unit area flow volumes applied to the ungauged single drainage 292 channel subwatersheds. Because stream discharge measurements may be accurate to within 5.0% 293 for continuous monitoring of river stage (Winter, 1981; Herschy, 1973), an accuracy of 5.0% for 294 each daily discharge estimate was assumed for each of the three rivers. An uncertainty of 9.0% 295 was applied to *R*_{sinale chan} based on the maximum difference observed by Devito and Dillon 296 (1993) for this type of extrapolation. The uncertainty for each subwatershed area (other than the lake itself) was assumed to be $\pm 1 \text{ km}^2$. The total uncertainty for the net river discharge (δR) was 297 298 calculated as shown in Table 1.

299 All river and direct runoff flow volumes were normalized by the (average) lake area (155 km²; OIVA/HERTTA, 10 Aug 2010). The variation in the area of the lake due to changes in lake 300 301 stage was assessed using ArcMap in order to estimate the uncertainty related to the chosen value. 302 Contour maps created from interpolation of topographic (MML, 2009c) and bathymetric (MML, 303 2009b) elevation data were used to calculate the maximum and minimum lake area based on the 304 range of lake stages observed between 1960 and 2010. The areas varied by less than 1.6% (2.5 km²) over that time (Wiebe, 2012). This value of uncertainty for the lake area was included in 305 306 calculations of both δR and δDR (Table 1).

307 **3.3.6. Direct Runoff**

The direct runoff (*DR*) for the six subwatersheds of the lake where non-channelized flow occurs (Figure 1, "Direct Runoff" subwatersheds) was estimated in two ways: 1) the use of a runoff coefficient method to estimate runoff as a percentage of precipitation, and 2) the

extrapolation of direct runoff values obtained by hydrograph separation analysis of a gaugedriver watershed within the lake's watershed.

313 Runoff Coefficient Method

314 Several runoff coefficient methods were reviewed for use in this study (e.g., Natural 315 Resources Conservation Service, 2004; Motz et al., 2001; Sacks et al., 1998; Dames and Moore, 316 1992; and Barazzuoli et al., 1989). To our knowledge, none have been used for estimating direct 317 runoff for large lakes. The CN method (Natural Resources Conservation Service, 2004) is a well-318 known method and more commonly used in other scenarios; however, it could not be used 319 because the soils of the Lake Pyhäjärvi watershed had not been classified according to the U.S. 320 Natural Resources Conservation Service format. The method best suited to the data available at 321 Lake Pyhäjärvi was the runoff coefficient map method developed by Kennessey (1930) and 322 modified by Barazzuoli et al. (1989). The method calculates average annual direct runoff using 323 the areal coverage of subcategories in three physiographic themes or "components" (surface soil 324 permeability, vegetation types, and slope angles) for a watershed over a given time period. The 325 components are summed to obtain a fraction of the precipitation that is direct runoff. ArcMap 326 10.0 was used to estimate the proportional coverage areas for the various categories of the 327 method (Table 2), using surface geology (GTK, 2008), land cover (SYKE, 2004), and elevation 328 (MML, 2009c) datasets with raster grid cells 25 m by 25 m in size. The appropriate set of 329 coefficients for the method was selected based on the index of aridity calculated for the 330 Jokioinen meteorological station using monthly and annual, daily-derived temperature and 331 precipitation averages from Pirinen et al. (2012). The coefficient map method was applied to the 332 six non-channelized direct runoff subwatersheds and also to the two gauged Yläneenjoki and 333 Pyhäjoki River watersheds so that the results also could be directly compared to the hydrograph

separation results for those two watersheds. Yearly precipitation values for these regions wereestimated with the interpolation procedure described above for *P*.

Assigning errors to a runoff coefficient map method is problematic because to our knowledge no rigorous evaluations of its absolute effectiveness have been performed. The most applicable error estimate found in the literature was an absolute uncertainty of \pm 0.16, which was the average error obtained from a study on the differences between observed and literature values of event runoff coefficients (Dhakal et al., 2012), and it was used in calculating δDR in Table 1. The uncertainty estimates for δP (described above) were also used for δDR calculations.

342

Hydrograph Separation Method

343 This method of determining direct runoff to the lake is based on the concept of 344 determining the direct runoff and groundwater flow components for gauged rivers within the 345 lake's watershed using hydrograph separation techniques and then applying the values to other 346 (non-channelized) areas of the lake watershed that have similar physiographic characteristics. 347 Despite the straight forward and intuitive nature of such an approach and the fact that 348 hydrograph separation techniques have continued to improve in past decades, it appears only 349 Newbury and Beaty (1980) and Schindler et al. (1976) have used a hydrograph separation 350 approach to extrapolate direct runoff from gauged subwatersheds of a lake to those with only 351 non-channelized flow.

In this study, the United States Geological Survey's PART automated hydrograph separation method was used (Rutledge, 2007). Several other techniques were considered but not used (e.g., HYSEP [Sloto and Crouse, 1996], UKIH [Piggott et al., 2005], BFLOW [Arnold and Allen, 1999], Eckhardt [Eckhardt, 2005]). The PART and HYSEP methods performed the best on average in an evaluation by Partington et al. (2012) and provided similar results in a study by

357 Eckhardt (2008). PART was selected because of its ease of use and because it is more commonly 358 used. The data required for the PART program included: daily streamflow measurements 359 (obtained from OIVA/HERTTA, 23 Sep 2010); the drainage area for the gauged region of each 360 river (OIVA/HERTTA, 23 Sep 2010); and the starting and ending years for the data sets. 361 Hydrograph records were processed using the following program settings: a threshold of 0.1 log 362 cycles per day for the daily decline in streamflow (Rutledge, 1998) and a value of N of N-1. N is 363 the number of days (as an integer) of impact that a rainfall event has after the peak flow when interflow and surface runoff components are significant (i.e., where N=A^{0.2} and A is watershed 364 365 area in square miles). The N-1 value was selected because it provided more accurate results for 366 similar types of rainfall and hydrographs in a comprehensive evaluation of hydrograph 367 separation methods performed by Partington et al. (2012). The PART method sums up daily calculated groundwater contribution values for the year and reports a base flow index (BFI) for 368 369 the river, which is the fraction (i.e., 0 to 1.0) of total river flow that is groundwater for the year. 370 The remainder of the fraction represents the streamflow that is contributed by direct runoff. 371 Direct precipitation onto, and evaporation off of, the river surface are considered negligible.

372 Annual direct runoff estimates were obtained for the Pyhäjoki River from 1972 to 2009 373 and for the Yläneenjoki River from 1971 to 2009. The Yläneenjoki River results were used to 374 estimate direct runoff in the six non-channelized direct runoff subwatersheds because the 375 surficial geology of the subwatersheds adjacent to the lake were finer grained and a better match 376 to the Yläneenjoki River watershed than to the coarser grained deposits of the Pyhäjoki River 377 watershed (Figure 2; Table 2). The percentage of land area covered by bedrock, till, and clay in 378 the Yläneenjoki watershed (76%) was slightly larger than the area for the direct runoff 379 subwatersheds (71%), while the Pyhäjoki coverage area (36%) was much smaller. The

percentage area covered by eskers, glaciofluvial materials, sands, and gravels in the Yläneenjoki
watershed (7%) was smaller than the percentage area in the direct runoff subwatersheds (15%),
while the Pyhäjoki coverage area (49%) was much larger. Direct runoff to the lake was estimated
by multiplying the total Yläneenjoki River flow per unit gauged area by the area of the direct
runoff subwatersheds and then by the direct runoff fraction (1 – BFI) obtained for the
Yläneenjoki River using PART for the corresponding time period.

386 A main uncertainty associated with using the PART method centres around whether the 387 automated graphical hydrograph interpolation method actually results in an accurate 388 quantification of the true base flow. Partington et al. (2012) assessed the absolute accuracy of 389 hydrograph separation techniques by simulating an artificial watershed and single precipitation 390 events using HydroGeoSphere (Therrien et al., 2010), but the results could not be reliably scaled 391 up to estimate the uncertainty for an entire year in our study. Therefore, lacking a comparison 392 between PART and a true value of baseflow, the uncertainty for the BFI estimated by PART was 393 derived from a study by Sanford et al. (2012), who compared results of PART with a hydrograph 394 separation technique employing continuously measured specific conductivity values in rivers in 395 Virginia over 18 months. Sanford et al. (2012) assessed two streams having average topographic 396 slopes similar in magnitude to those in the Yläneenjoki and direct runoff subwatersheds (< 4%). 397 These two streams yielded absolute percentage differences between the chemical hydrograph 398 separation technique and PART (i.e., |BFI_{chem} – BFI_{PART} / BFI_{PART}) of 8.9% and 7.5%, 399 respectively. The average of these two values (8.2%) was assumed to be representative of the 400 relative uncertainty for BFI values from the PART method. For the extrapolation of values from 401 one subwatershed to another, the 9.0% uncertainty from Devito and Dillon (1993) was again 402 applied (Table 1).

403 **3.4 Quantifying Known Groundwater Discharges**

404 Although the calculation of *G* using the water balance equation includes all groundwater 405 inputs, groundwater discharge to the lake was calculated for an area of the shoreline where 406 significant amounts of groundwater discharge were known to occur (Rautio, 2009; Korkka-407 Niemi et al., 2011; Rautio and Korkka-Niemi, 2011) and used as an independent value to 408 compare to *G*. The groundwater discharge into the lake through the Honkala Aquifer in the 409 Kuivalahti-Säkylä esker was estimated by Wiebe (2012) using Darcy's Law and a hydraulic conductivity value of $K = 1 \times 10^{-3}$ m/s ± one order of magnitude, which was chosen to represent 410 411 flow in the coarse-grained esker core (Artimo, 2002). The cross-sections for the calculation are 412 shown on Figure 3. Discharge estimates were normalized by the lake area, and an uncertainty 413 estimate was developed according to the general procedures in Table 1. This approach to 414 estimating groundwater discharge into sections of shorelines of large lakes is not new; Singer 415 (1974) used the same approach to estimate flow into Lake Ontario in Canada.

416

417 **4. Results**

The water balance components and net groundwater (*G*) values calculated for each year using the runoff coefficient map method of obtaining direct runoff are shown in Figure 4. Nearequilibrium (i.e., near zero values) or net groundwater recharge occurs during the 1970s, a small amount of net discharge occurs during the 1980s, and mostly net groundwater recharge occurs from the 1990s until to the end of the study period. Fifteen of the 38 water years in the water balance appear to have net groundwater discharge conditions (of which only three are larger than the calculated uncertainty), while 23 of the 38 water years appear to have net groundwater

425 recharge conditions (of which nine are larger than their associated uncertainty). Overall, the 426 magnitude of *G* was less than the uncertainty during 26 water years. Table 3 summarizes the 427 average value for each component of the water balance for the entire study period. The average 428 total inflow and outflow for this water balance were 1481 mm and 1414 mm per year (not 429 including the groundwater component). The average value of *G* was -24 mm (-1.7% of average 430 total outflow) and indicates average net groundwater recharge conditions for the lake. During the 431 study period, the magnitude of *G* ranged from -268 mm to 268 mm with a standard deviation of 432 117 mm. The average uncertainty was 119 mm. The estimates of direct runoff to the lake ranged 433 from 83 mm to 167 mm per year and averaged 130 mm during the study period.

434 The water balance components and net groundwater values calculated for each year using 435 the PART hydrograph separation method to estimate direct runoff are shown in Figure 5. The 436 figure shows near equilibrium groundwater discharge conditions for the 1970s, groundwater 437 discharge conditions for the 1980s, and near equilibrium conditions for the 1990s onward. The 438 average value of *G* was +43 mm and indicates overall net groundwater discharge conditions. 439 This *G* represents about 3.0% of the average total inflow for the lake (i.e., 1414 mm, not 440 including the groundwater component). The magnitude of *G* ranged from -197 mm to 284 mm 441 with an average standard deviation of 118 mm during the study period. Twenty-six of the 38 442 water years appear to have net groundwater discharge conditions, and the magnitude of *G* was 443 greater than the calculated uncertainty for 12 of these years (Figure 6). The average uncertainty 444 was 103 mm. The estimates of direct runoff to the lake ranged from 14 mm to 140 mm and 445 averaged 63 mm during the study period.

Table 3 summarizes the average uncertainties associated with each component of thewater balance equation over the 38 year study period for both the coefficient map and

448 hydrograph separation methods for determining direct runoff. For both methods the component 449 having the largest average absolute uncertainty was evaporation (67.5 mm). The second largest 450 uncertainty of the PART derived method was precipitation (61.2 mm), while for the coefficient 451 map method both the direct runoff term and precipitation had the second largest uncertainty 452 (each equal to 61.2 mm). For the hydrograph separation method, components with the next 453 largest average absolute uncertainties were the change in lake storage, net river flow, and then 454 the direct runoff term (13 mm). As noted above, the uncertainty that accumulated in the net 455 groundwater term was 119 mm for the coefficient map method and 103 mm for the hydrograph 456 separation method. The difference between the two uncertainty values associated with *G* is a 457 direct result of the accuracy of the direct runoff component because all the other components 458 were calculated in the same way for both methods.

459 Relative and absolute uncertainties associated with each water balance component 460 differed from year to year during the study, depending on the component. For the hydrograph 461 separation method the ranges in uncertainties for the 38 year period were as follows. The 462 absolute uncertainty of the net groundwater flow component ranged from 77 mm to 135 mm and 463 the relative uncertainty ranged from 37% to 3800% (values greater than 100% mean the 464 uncertainty is greater than the value of the component). For evaporation the absolute uncertainty 465 ranged between 50 mm and 85 mm while the relative uncertainty was fixed at 15% (as described 466 earlier). The absolute uncertainty for precipitation ranged between 24 mm and 112 mm, and 467 relative uncertainty ranged between 5.2% and 20%. The relative net river inflow uncertainty 468 ranged between 4.1% and 89%, and the relative direct runoff uncertainty ranged from 13% to 36%. The lake water level absolute uncertainty (36 mm) was constant as described earlier, and 469

the relative uncertainty ranged between 9.5% and 361%. The estimated relative uncertainty onthe pumping withdrawals at Lohiluoma was 11%.

The groundwater baseflow and direct runoff for the Yläneenjoki and Pyhäjoki Rivers also 472 473 changed from year to year (Figure 7). The average baseflow index (BFI) obtained from the 474 PART hydrograph separation model for the study period was 0.68 for the Yläneenjoki River and 475 0.84 for the Pyhäjoki River, while the standard deviations for the two were 0.069 and 0.043, 476 respectively. The average indirect groundwater contributions from these rivers to the lake were 477 327 mm and 123 mm for the Yläneenjoki and Pyhäjoki, respectively. Overall, the average 478 indirect groundwater contribution to Lake Pyhäjärvi was at least 454 mm (or about +32% of 479 average total inflow when compared to the PART derived water balance). Based on these 480 average BFI values, the corresponding average direct runoff values for the Yläneenjoki and 481 Pyhäjoki Rivers during the study period were 32% and 16% of the river flow, respectively. 482 Figure 7 also shows how the values of direct runoff per unit area for each of the rivers' 483 watersheds varied during the study period. The finer grained deposits of the Yläneenjoki River 484 watershed resulted in direct runoff values that ranged from 23 mm to 228 mm per unit gauged 485 area of its watershed per year (with an average value of 103 mm per unit gauged area of its 486 watershed per year), and were, on an annual basis, consistently 3.3% to 29% higher than those 487 for the Pyhäjoki River watershed.

The coefficient map method of determining direct runoff to the lake during the study period resulted in values that were typically higher than those determined using the PART hydrograph separation method. In order to perform a direct comparison of the coefficient map and PART methods of determining direct runoff, both methods were applied to both of the gauged river watersheds. The average surface flow estimated by PART constituted 7.8% of the

493 precipitation (estimated using the isohyetal method) in the gauged region of the Pyhäjoki River 494 watershed, and 16% of the isohyetally derived precipitation estimate in the gauged region of the 495 Yläneenjoki River watershed. The corresponding percentages for the coefficient map method 496 were 29% and 38% for the gauged regions of the Pyhäjoki River and Yläneenjoki River 497 watersheds, respectively (Table 2). The runoff coefficient map estimates were about 2.4 to 3.7 498 times higher than those estimated by the PART method. Higher direct runoff values result in 499 lower amounts of precipitation entering the groundwater.

500 The amount of groundwater estimated to directly enter the lake through the Honkala 501 Aquifer in the Kuivalahti-Säkylä esker appears to be significant and relatively constant but is 502 subject to significant uncertainty. The groundwater discharge from the esker was estimated to be 503 22 mm or about 1.6% of the average total inflow for the PART derived water balance, with an 504 uncertainty of ± one order of magnitude. The uncertainty in the hydraulic conductivity value of 505 the geological materials was responsible for essentially all the uncertainty in this Darcian flow estimate. However, the average estimated Darcy flux of 6.81×10^{-6} m/s (per unit cross section 506 area of the aquifer) is consistent with and in the 10^{-7} to 10^{-5} m/s range for groundwater discharge 507 508 into the lake measured by Rautio (2009) using seepage meters where the esker and aquifer 509 intersect the shoreline. The amount of groundwater entering the lake from the aquifer each year 510 is likely relatively constant because the water levels (and hydraulic gradients) in the Honkala 511 Aquifer are relatively constant (Artimo, 2002), as are the regulated lake levels that vary within a 512 1 m range. The average groundwater discharge from the aquifer appears equal to approximately 513 half of the +43 mm average net groundwater component for the entire lake estimated using the 514 PART method.

515

516 **5. Discussion**

517 An accurate and scientifically meaningful water balance for a large lake requires: 518 collecting a considerable amount of data, implementing successful upscaling schemes, 519 employing techniques for estimating components that are not easily measured, and understanding 520 the uncertainty related to both the measurement methods and their spatial and temporal 521 extrapolation methods. The lack of data or lack of good quality data, or an inability to properly 522 quantify or reduce errors that accumulate in the calculation method has often led to water 523 balances that do not even attempt to quantify net groundwater flow for large lakes (e.g., Kuusisto 524 (1975) and Järvinen (1978) for Lake Pyhäjärvi). The main factor that prevents calculation of the 525 net groundwater flow component is the inability to accurately separate out and quantify direct 526 runoff contributions from a term that lumps all groundwater flow with non-channelized overland 527 flow and interflow from subwatersheds with no streams. A second and almost equally large 528 problem is that unless all uncertainties in the water balance equation components are accurately 529 quantified, one will not know if the calculated net groundwater value is real (i.e., larger than the 530 accumulated error) or not. This study of Lake Pyhäjärvi appears to have resolved these two 531 problems by using the PART hydrograph separation method to estimate and minimize the 532 uncertainty related to the direct runoff to the lake, and by employing the rigorous uncertainty 533 analysis summarized in Table 1. The key to the success of this study was the opportunity to use 534 stream flow gauging data for a river within the lake's watershed and then apply the results to 535 non-channelized parts of the watershed.

The two main advantages to using the PART hydrograph separation method to estimate direct runoff for the lake is the site specific representativeness of the technique and the relatively low uncertainty associated with the method. Unlike the coefficient mapping technique, the PART

539 method develops an empirical relationship between direct runoff and the actual rainfall events 540 (magnitudes and intensities), antecedent conditions (i.e., moisture contents of soils), geology, 541 vegetation types, topographic slopes, and groundwater flow processes actually occurring in the 542 lake's watershed. In contrast, the runoff coefficient map method cannot appropriately deal with 543 overland drainage to a low-lying area where water infiltrates or evaporates rather than flowing to 544 the lake, and it does not account for rainfall intensity or antecedent soil conditions. Furthermore, 545 the runoff coefficient map method is unable to produce different percentages of runoff versus 546 rainfall for different water years and is unable to adapt to a climate having precipitation that 547 varies over a range of several hundred millimetres per water year. These deficiencies in the 548 coefficient mapping method resulted in: 1) estimates of direct runoff that were on average 2.4 to 549 3.7 times higher than those from the PART method, 2) an underestimation of direct groundwater contributions to the lake, and 3) the conclusion that on average during the 38 years the lake was 550 551 losing surface water to (i.e., recharging) the groundwater at a rate (*G*) of at least -24 mm despite 552 the fact that none of the field investigations have detected significant losing areas within the lake. 553 Moreover, there do not appear to have been sufficient studies to definitively verify the accuracy 554 of the coefficient map method developed by Kennessey (1930) and modified by Barazzuoli et al. 555 (1989) or other similar methods on an annual basis. The CN method (Natural Resources 556 Conservation Service, 2004) attempts to account for issues such as the antecedent moisture 557 content of the soil and the threshold rainfall that will generate runoff, but it still does not 558 incorporate rainfall intensity. In contrast, the PART method for determining the direct runoff 559 provided a more realistic average groundwater discharge rate of +43 mm (overall a gaining lake 560 condition) and the uncertainty associated with the method is smaller.

561 The magnitude of uncertainties assigned to the PART method itself and the calculated 562 values of direct runoff appear to be reasonable and accurate, and the concept of assigning those 563 values to adjacent non-channelized watersheds appears to be valid. The relative uncertainty of \pm 564 8.2% from the comparison by Sanford et al. (2012) of PART results and a chemical hydrograph 565 separation method was the average value for two watersheds in Virginia with average 566 topographic slopes similar to those found in the Yläneenjoki River watershed. This uncertainty 567 value corresponds well to the estimated uncertainty derived from a controlled numerical 568 experiment by Partington et al. (2012). The average absolute difference of \pm 0.023 per event 569 from that study, when upscaled to the average number of similar events (12) in the Yläneenjoki 570 River per water year, yields an absolute uncertainty of \pm 8.0%. Further, annual absolute percent 571 differences between PART and the HYSEP-sliding interval or HYSEP-fixed interval programs 572 were also between 6.5 and 8.4% in a study by Risser et al. (2005). The concept of using PART to 573 calculate stream baseflow in a gauged watershed and extrapolating results to physically similar 574 watersheds was performed quite successfully in the Great Lakes watershed in Canada and the 575 USA (Neff et al., 2005). The Neff et al. (2005) empirical approach for extrapolating the PART 576 results to other watersheds was more sophisticated than that employed in this study and was 577 based on data from hundreds of gauging stations and demonstrated the validity of this type of 578 approach. In the Lake Pyhäjärvi watershed, extrapolating the PART derived direct runoff results 579 from the Yläneenjoki River watershed to the adjacent non-channelized watersheds was clearly 580 more appropriate than using the results for the Pyhäjoki River based on the geological 581 considerations. For lake watersheds that have a sufficient number of streams within them, it 582 should be possible to select the most appropriate ones for streamflow gauging and subsequent 583 extrapolation to non-channelized portions of the lake watershed.

584 The comprehensive and detailed uncertainty analysis performed in this study provided 585 the information necessary to confidently assess if small values of net groundwater discharge 586 were real or still too uncertain to be determined reliable, and this methodology can be applied to 587 other lake water balances. The uncertainty calculation methodology presented in this study 588 builds on earlier work on uncertainties in lake water balances by Winter (1981), Lee and 589 Swancar (1997), and Neff and Nicholas (2005). As shown in Table 1, most of the uncertainty 590 values for individual components of the water balance cannot be obtained simply from these 591 earlier publications or from a single literature value but instead must be calculated using 592 knowledge of the site specific techniques used to collect and calculate each water balance 593 component. The equations contained in Table 1 can be adapted and used at other sites, and Table 594 1 also provides specific values of uncertainty for components that have not previously been 595 quantified for the purpose of a lake water balance (e.g., the uncertainty associated with using 596 PART to calculate BFI values and estimate direct runoff). In this study, magnitudes of *G* as small 597 as 8% of the total inputs could be reliably determined (depending on the particular water year) 598 and the value of δG provides a meaningful upper boundary for what *G* can be for years when the 599 uncertainty is larger than *G*. Over the 38 year period, the PART derived water balance results 600 indicated that 17 values of G that were greater than δG ranged between 101 mm and 284 mm and 601 represented 8% to 19% of the total inflows (without groundwater) for those years. The *G* values 602 obtained in this study are significant enough to be measureable but still a very small part of the 603 overall water balance, whereas indirect groundwater discharge to the lake (via the rivers) is very 604 significant and on average accounted 454 mm or about 32% of the total inputs to the lake during 605 the study period.

606 A water balance should be conducted in addition or as an alternative to numerical 607 modelling of groundwater-surface water interactions involving large lakes. Because a water 608 balance method can involve quantification of the uncertainties on the various individual 609 components, it can clarify the reliability of the component estimates and present meaningful 610 error bars. Numerical models inherently struggle with accurately defining boundary conditions 611 and with appropriately representing the hydrogeological properties (often having several orders 612 of magnitude variability and uncertainty) and other characteristics of the site. It can be argued 613 that large numerical models based on sparse data sets may introduce more uncertainty and make 614 uncertainties unquantifiable with respect to estimates of *G* or other components because of the 615 large number of assumptions and wide range of possible values needed to populate such models 616 (e.g., hydraulic conductivities, unsaturated zone flow characteristics). Performing a water 617 balance (as shown here) is a necessary first step to providing the data, calibration targets, and 618 reality checks needed for numerical models to provide meaningful predictions.

619

620 6. Conclusions

621 This study of Lake Pyhäjärvi and its watershed demonstrated that minimizing and 622 carefully quantifying uncertainties in the components used in the lake's water balance 623 calculations is the key to determining meaningful estimates of net groundwater flow for a large 624 lake, especially if net groundwater contributions are a relatively small part of the water balance. 625 The estimate of the direct runoff component of the lake water balance was improved by using the 626 PART hydrograph separation derived estimates of runoff for a river within the lake's watershed 627 and then applying those values to the non-channelized areas of the lake's watershed. The key is 628 to use direct runoff estimates obtained from a river that has similar characteristics (e.g., geology,

topography, vegetation, antecedent rainfall conditions, and rainfall intensities) to the nonchannelized areas, which was why the direct runoff values for the Yläneenjoki River (average BFI = 0.68) were used instead of those derived for Pyhäjoki River (average BFI = 0.84). The main advantage of this approach to estimating direct runoff to the lake is that it represents an empirical relationship that is calibrated to an actual portion of the lake's watershed and climatic conditions, unlike other coefficient runoff methods or numerical models that rely on relationships developed elsewhere and for very different watershed conditions.

636 The average net groundwater flow and the indirect groundwater discharge for the lake 637 were quantified in this study. The average net groundwater flow into Lake Pyhäjärvi over the 38 638 water years between October 1971 and September 2009 was calculated to be +43 mm (3.0% of 639 average total inflow) using the PART derived direct runoff values (average: 63 mm) for the 640 Yläneenjoki River. The uncertainty analysis showed that the magnitude of the net groundwater 641 flow was greater than the overall uncertainty in 17 out of 38 water years. A positive net 642 groundwater flow value represents the minimum possible value of direct groundwater discharge 643 to the lake (i.e., when groundwater recharge is zero), and if parts of the lake are also losing 644 surface water to groundwater (i.e., recharging groundwater), the direct discharges could be 645 proportionally larger. It is not known if any areas of the lake are recharging the groundwater, but 646 previous field investigations suggest that the lake is gaining groundwater rather than losing 647 surface water (Rautio, 2009; Korkka-Niemi et al., 2011; Rautio and Korkka-Niemi, 2011). A 648 significant amount of the direct groundwater discharge into the lake may occur through the 649 Honkala Aquifer in the Kuivalahti-Säkylä esker, which was estimated using Darcy's Law to be 650 22 mm (about 1.6% of the average total inflow for the PART derived water balance), but that 651 estimate has an uncertainty of \pm one order of magnitude. Independent, field-based measurements

652 of groundwater discharge provide an important check on the magnitude of the net groundwater 653 flow values, and if larger than that value, they can be used to infer that part of the lake must be 654 recharging the aquifer. Although direct groundwater discharges actually may be much larger 655 than the net value calculated, it is clear that indirect discharges of groundwater to the lake play a 656 major role in the water balance. The total average indirect groundwater contribution to the lake 657 from the Yläneenjoki and Pyhäjoki River discharges was 454 mm (+32% of average total 658 inflow), which indicates that the groundwater entering the rivers can have a large influence on 659 the quantity and quality of the water in the lake.

660 The techniques used in this study are applicable to other large lakes with inflowing 661 streams and rivers and may allow small net groundwater flows to be reliably quantified in 662 situations that might otherwise be unquantifiable or cause values to be completely lost in large 663 uncertainties.

664

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671

672 **References**

673	Arnold, J.G., and Allen, P.M., 1999. Validation of automated methods for estimating baseflow
674	and groundwater recharge from stream flow records. J. Am. Water Resour. Assoc. 35 (2), 411–
675	424.

Artimo, A., 1998. Säkylän Honkalan pohjavesialueen likaantumistapaus: Geologiset tutkimukset.
Lounais-Suomen ympäristökeskus. Unpublished Report. [In Finnish].

679

- Artimo, A., 2002. Application of flow and transport models to the polluted Honkala aquifer,
- 681 Säkylä, Finland. Boreal Env. Res. 7, 161–172. ISSN 1239-6095.

682

Ayenew, T., and Gebreegziabher, Y., 2006. Application of a spreadsheet hydrological model for
computing the long-term water balance of Lake Awassa, Ethiopia. Hydrolog. Sci. J. 51(3), 418–
431. doi:10.1623/hysj.51.3.418.

686

Baird, A.J., and Wilby, R.L., 1999. Eco-Hydrology: Plants and Water in Terrestrial and Aquatic
Environments. Rutledge Press, New York.

- 690 Barazzuoli, P., Izzo, S., Menicori, P., Micheluccini, M., and Salleolini, M., 1989. A New
- 691 Practical Aid to Regional Hydrogeologic Planning: The Runoff Coefficient Map. Environ.
- 692 Manage. 13(5), 613–622.
- 693

Bruce, J.P., Cunningham, W., Freeze, A., Gillham, R., Gordon, S., Holysh, S., et al., 2009. The
Sustainable Management of Groundwater in Canada. Ottawa, Canada: The Canadian Council of
Canadian Academies.
Cartwright, K., Hunt, C.S., Hughes, G.M., and Brower, R.D., 1979. Hydraulic potential in Lake
Michigan bottom sediments: Journal of Hydrology, 43 (1/4), 67–78. doi:10.1016/0022-
1694(79)90165-3.
Cherkauer, D.S., and Nader, D.C., 1989. Distribution of groundwater seepage to large surface-
water bodies: The effect of hydraulic heterogeneities. J. Hydrol. 109 (1-2), 151–165.
doi:10.1016/0022-1694(89)90012-7.
Dames and Moore, 1992. City of Lakeland comprehensive stormwater management and lake
pollution study, v. I: Consultant's report submitted to City of Lakeland, FL.

- 709 Demlie, M., Ayenew, T., and Wohnlich, S., 2007. Comprehensive hydrological and
- 710 hydrogeological study of topographically closed lakes in highland Ethiopia: The case of Hayq
- 711 and Ardibo. J. Hydrol. 339, 145–158. doi:10.1016/j.jhydrol.2007.03.012.

- 713 Devito, K.J., and Dillon, P.J., 1993. Errors in estimating stream discharge in small headwater
- 714 catchments: Influence on interpretation of catchment yields and input output budget estimates.

- 715 Tech. Rep. 1993. Ont. Minist. Environ., Dorset, Ontario, Canada.
- 716 https://archive.org/details/errorsinestimati00deviuoft. Accessed 11 Mar 2015.

- 718 Dhakal, N., Fang, X., Cleveland, T.G., Thompson, D.B., Asquith, W.H., and Marzen, L.J., 2012.
- 719 Estimation of volumetric runoff coefficients for Texas watersheds using land-use and rainfall-
- runoff data. J. Irrig. Drain. Eng. 138 (1), 43–54. doi:10.1061/(ASCE)IR.1943-4774.0000368.

721

722 Dingman, S.L., 1994. Physical Hydrology. Prentice Hall, Englewood Cliffs, NJ.

723

- 724 Dubrovsky, N.M., Burow, K.R., Clark, G.M., Gronberg, J.A.M., Hamilton, P.A., Hitt, K.J.,
- 725 Mueller, D.K., Munn, M.D., Nolan, B.T., Puckett, L.J., Rupert, M.G., Short, T.M., Spahr, N.E.,
- 726 Sprague, L.A., and Wilber, W.G., 2010. The Quality of Our Nation's Water Nutrients in the
- Nation's Streams and Groundwater, 1992-2004. U.S. Geological Survey Circ. 1350, 174p.
- 728 http://pubs.usgs.gov/circ/1350/pdf/circ1350.pdf. Accessed 1 Dec 2011.

729

- Eckhardt, K., 2005. How to construct recursive digital filters for baseflow separation. Hydrol.
- 731 Processes 19, 507–515. doi:10.1002/hyp.5675.

732

- 733 Eckhardt, K., 2008. A comparison of base flow indices, which were calculated with seven
- different base flow separation methods. J. Hydrol. 352, 168–173.
- 735 doi:10.1016/j.jhydrol.2008.01.005.

737	Eronen, M., Heikkinen, O., and Tikkanen, M., 1982. Holocene development and present
738	hydrology of Lake Pyhäjärvi in Satakunta, southwestern Finland. Fennia 160 (2), 195–223.
739	
740	ESRI, 2009. Europe. 9.3.1 Media Kit. Available: University of Waterloo Map and Design
741	Library. GIS digital mapping data. Accessed 15 Nov 2011.
742	
743	ESRI, 2010. ArcMap 10.0. http://www.esri.com/.
744	
745	Feinstein, D.T., Hunt, R.J., and Reeves, H.W., 2010. Regional groundwater-flow model of the
746	Lake Michigan Basin in support of Great Lakes Basin water availability and use studies. U.S.
747	Geological Survey Sci. Inv. Rep. 2010–5109. 379 p.
748	
749	Finnish Meteorological Institute (FMI), 2011. Point precipitation measurements. Available on
750	request. http://en.ilmatieteenlaitos.fi/.
751	
752	Fruh, E.G., 1967. The Overall Picture of Eutrophication. J. Water Pollution Control Fed. 39 (9),
753	1449–1463.
754	
755	Golden Software, Inc., 2002. Surfer User's Guide.
756	http://www.wi.zut.edu.pl/gis/Surfer_8_Guide.pdf. Accessed 1 Feb 2012.

758	GTK, 2008. Surface soil geology [maaperä] 1:20 000, Kaista 1 (Sheets 1133 and 1134) and 2
759	(Sheets 2111 and 2212.). GIS digital mapping data. Online distribution.
760	
761	Häkkinen, K., 1996. Suomalaisten esihistoria kielitieteen valossa. Suomalaisen Kirjallisuuden
762	Seura, Helsinki. 224p. [In Finnish].
763	
764	Harvey, F.E., Rudolph, D.L., and Frape, S.K., 1997. Measurement of hydraulic properties in
765	deep lake sediments using a tethered pore -pressure probe: Application in the Hamilton Harbour,
766	Western Lake Ontario. Water Resour. Res. 33 (8), 1917–1928.
767	
768	Harvey, F.E., Rudolph, D.L., and Frape, S.K., 2000. Estimating ground water flux into large
769	lakes: Application in the Hamilton Harbor, western Lake Ontario: Ground Water, 38 (4), 550–
770	565.
771	
772	Harbeck, G.E., Jr., Kennon, F.W., Kohler, M.A., et al., 1954. Water Loss Investigations: Lake
773	Hefner Studies, Technical Report. U.S. Geological Survey Prof. Paper 269.
774	
775	Hayashi, M., and Rosenberry, D.O., 2002. Effects of ground water exchange on the hydrology
776	and ecology of surface water. Ground Water 40 (3), 309–316. doi:10.1111/j.1745-
777	6584.2002.tb02659.x.

779

780	Koblenz Symposium on Hydrometry (1970). IAHS Publ. 99, 109–131.
781	
782	Hoaglund, J.R., Huffman, G.C., and Grannemann, N.G., 2002. Michigan Basin regional ground
783	water flow discharge to three Great Lakes. Ground Water, 40 (4), 390–406. doi:10.1111/j.1745-
784	6584.2002.tb02518.x.
785	
786	Holtschlag, D.J., and Nicholas, J.R., 1998. Indirect Groundwater Discharge to the Great Lakes.
787	U.S. Geological Survey, Open-File Rep. 98-579.
788	
789	Hounam, C.E., 1973. Comparison between Pan and Lake Evaporation. World Meteorological
790	Organization, Technical Note No. 126, 52p.
791	
792	Hyvärinen, V., Järvinen, J., and Tuominen, T., 1973. Water balance of Lakes Pyhäjärvi and
793	Pääjärvi. Proceedings of the Helsinki Symposium. IAHS Publ. 109, 276–283.
794	
795	Järvinen, J., 1978. Estimating Lake Evaporation with Floating Evaporimeters and with Water
796	Budget. Nord. Hydrol. 9, 121–130.
797	
798	Kennessey, B., 1930. Lefolyasi téniezok és retenciok. Vizugy, Koziemények, Hungary.
	36

Herschy, R.W., 1973. The magnitude of errors at flow measurement stations. Proceedings of the

800	Kirkkala, T., 2014. Long-term nutrient load management and lake restoration: Case of Säkylän
801	Pyhäjärvi (SW Finland). Ph.D. dissertation, University of Turku, Turku, Finl., 55p.
802	http://www.doria.fi/bitstream/handle/10024/94354/AnnalesAII286Kirkkala.pdf?sequence=2.
803	Accessed 24 Apr 2015.
804	
805	Kohler, M.A., Nordenson, T.J., and Baker, D.R., 1959. Evaporation maps for the United States.
806	U.S. Weather Bureau. Tech. Paper No. 37.
807	
808	Korkka-Niemi, K., Rautio, A., Niemistö, P., and Karhu, J.A., 2011. Hydrogeochemical and
809	isotopic indications of groundwater-surface water interactions at Lake Pyhäjärvi, SW Finland.
810	GQ10: Groundwater Quality Management in a Rapidly Changing World. Proceedings of the 7th
811	International Groundwater Quality Conference, Zurich, Switzerland, 13-18 Jun 2010. IAHS
812	Publ. 342, 423–426.
813	
814	Kuusisto, E., 1975. Säkylän Pyhäjärven vesitase ja säännöstely (English summary: The water
815	balance and regulation of Lake Pyhäjärvi). Vesihallitus - National Board of Waters, Finl.
816	Publications of the Water Research Institute 11. 86p. [In Finnish with English Summary.]
817	
818	LaBaugh, J.W., Rosenberry, D.O., and Winter T.C., 1995. Groundwater contribution to the water
819	and chemical budgets of Williams Lake, Minnesota, 1980–1991. Can. J. Fish. Aquat. Sci. 1995,
820	52 (4), 754–767.
	37

822	Lee, T.M., and Swancar, A., 1997. Influence of Evaporation, Ground Water, and Uncertainty in
823	the Hydrologic Budget of Lake Lucerne, a Seepage Lake in Polk County, Florida. U.S.
824	Geological Survey Wat. Sup. Paper 2439.
825	
826	Lenters, J.D., 2004. Trends in the Lake Superior Water Budget since 1948: A Weakening
827	Seasonal Cycle. J. Great Lakes Res. 30 (Supplement 1), 20–40.
828	
829	Linsley, R.K., Kohler, M.A., and Paulhus, J.L.H., 1958. Hydrology for Engineers. McGraw-W,
830	New York, NY.
831	
832	Luoto, M., 2000. Spatial analysis of landscape ecological characteristics of five agricultural areas
833	in Finland by GIS. Fennia 178 (1), 15–54.
834	
835	Menking, K.M., Syen, K.H., Anderson, R.Y., Shafike, N.G., and Arnold, J.G., 2003. Model
836	estimates of runoff in the closed, semiarid Estancia basin, central New Mexico, USA. Hydrolog.
837	Sci. J. 48 (6), 953–970. doi:10.1623/hysj.48.6.953.51424.
838	
839	MML, 2009a. Base maps [peruskartta], UL3424L_RK1_1.tif, UL3424R_RK1_1.tif,
840	UL3442L_RK1_1.tif, UM3313L_RK1_1.tif, UM3313R_RK1_1.tif, UM3331L_RK1_1.tif,

841	UM3312R_RK1_1.tif, UM3314L_RK1_1.tif, UM3314R_RK1_1.tif, and UM3323L_RK1_1.tif.
842	GIS digital mapping data.
843	
844	MML, 2009b. Bathymetric elevations, skunta_syvpiste.shp. N60. GIS digital mapping data.
845	
846	MML, 2009c. Topographic elevations [korkeusmalli], 25m point spacing, 20n_dem2.xyz and
847	20m_dem2.xyz. N60. GIS digital mapping data.
848	
849	Motz, L.H., Sousa, G.D., and Annable, M.D., 2001. Water budget and vertical conductance for
850	Lowry (Sand Hill) Lake in north-central Florida, USA. J. Hydrol. 250, 134–148.
851	
852	Mylopoulos, N., Mylopoulos, Y., Veranis, N., and Tolikas, D., 2007. Groundwater modeling and
853	management in a complex lake-aquifer system. Water Resour. Manage. 21, 469–494.
854	doi:10.1007/s11269-006-9025-3.
855	
856	Natural Resources Conservation Service, 2004. Estimation of direct runoff from storm rainfall,
857	in: Part 630 Hydrology, National Engineering Handbook. National Soil Conservation Service
858	(Formerly Soil Conservation Service), U.S. Dept. of Agriculture, Washington, DC (Chapters 9
859	and 10).
860	

- 861 Neff, B.P., and Killian, J.R., 2003. The Great Lakes water balance: data availability and
- annotated bibliography of selected references. U.S. Geological Survey Water-Resour. Inv. Rep.02-4296. 37p.

- 865 Neff, B.P., and Nicholas, J.R., 2005. Uncertainty in the Great Lakes Water Balance. Date Posted:
- 866 23 November 2005. U.S. Geological Survey Sci. Inv. Rep. 2004-5100, 42p.
- http://pubs.water.usgs.gov/sir2004-5100/. Accessed 5 Feb 2014.

868

- 869 Neff, B.P., Day, S.M., Piggott, A.R., and Fuller, L.M., 2005. Base Flow in the Great Lakes
- 870 Basin. U.S. Geological Survey Sci. Inv. Rep. 2005-5217. 23 p.
- 871 http://pubs.water.usgs.gov/sir2005-5217.
- 872
- 873 Newbury, R.W., and Beaty, K.G., 1980. Water renewal efficiency of watershed and lake
- 874 combinations in the ELA region of the Precambrian Shield. Can. J. Fish. Aquat. Sci. 37, 335–

875 341.

876

- 877 OIVA/HERTTA, 2010 2011. Finnish Environment Administration Data.
- 878 http://www.ymparisto.fi/OIVA.

880	Panday, S., and Huyakorn, P.S., 2004. A fully coupled physically-based spatially-distributed
881	model for evaluating surface/subsurface flow. Adv. Water Resour. 27, 361–382.
882	doi:10.1007/s11269-006-9025-3.
883	

- Partington, D., Brunner, P., Simmons, C.T., Werner, A.D., Therrien, R., Maier, H.R., and Dandy,
- 885 G.C., 2012. Evaluation of outputs from automated baseflow separation methods against
- simulated baseflow from a physically based, surface water-groundwater flow model. J. Hydrol.
- 458-459, 28–39. doi:10.1016/j.jhydrol.2012.06.029.

Piggott, A.R., Moin, S., and Southam, C., 2005. A revised approach to the UKIH method for the
calculation of baseflow. Hydrol. Sci. J. 50, 911–920. doi:10.1623/hysj.2005.50.5.911.

891

- Pirinen, P., Henriikka, S., Aalto, J., Kaukoranta, J.-P., Karlsson, P., and Ruuhela, R., 2012.
- 893 Climatological statistics of Finland 1981-2010. Finnish Meteorological Institute, Helsinki, Finl.
- http://en.ilmatieteenlaitos.fi/normal-period-1981-2010. Accessed 19 June 2012.

895

Power, G., Brown, R. S., and Imhof, J.G., 1999. Groundwater and fish - insights from northern
North America. Hydrol. Process. 13, 401–422.

- 899 Quinn, F.H., and Guerra, B., 1986. Current perspectives on the Lake Erie Water Balance. J.
- 900 Great Lakes Res. 12 (2), 109–116.

901	
-----	--

902	Ramette, R.W., 1981. Chemical equilibrium and analysis. Addison-Wesley, Reading, MA.
903	
904	Rautio, A., 2009. Characterization of Groundwater-Lake Interactions at Lake Pyhäjärvi, SW
905	Finland. M.Sc. Thesis, University of Turku, Turku, Finl., 92p.
906	
907	Rautio, A., and Korkka-Niemi, K., 2011. Characterization of groundwater–lake water
908	interactions at Pyhäjärvi, a lake in SW Finland. Boreal Environ. Res. 16, 363–380. ISSN 1797-
909	2469.
910	
911	Reko, J., 2010. Personal communication from Environmental Manager, Municipality of Säkylä,
912	Finland.
913	
914	Risser, D.W., Gburek, W.J., and Folmar, G.J., 2005. Comparison of methods for estimating
915	ground-water recharge and base flow at a small watershed underlain by fractured bedrock in the
916	eastern United States. U. S. Geological Survey Sci. Inv. Rep. 2005-5038, 31p. Reston, VA.
917	http://pubs.usgs.gov/sir/2005/5038/pdf/sir2005-5038.pdf. Last updated: 17 Jan 2014. Accessed
918	24 Jun 2014.
919	
920	Rosenberry D.O., Striegl, R.G., Hudson, D.C., 2000. Plants as indicators of focused ground

921 water discharge to a northern Minnesota lake. Ground Water, 38 (2), 296–303.

923	Rutledge, A.T., 1998. Computer Programs for Describing the Recession of Ground-Water
924	Discharge and Estimating Mean Ground-Water Recharge and Discharge from Streamflow
925	Records – Update. U.S. Geological Survey Inv. Rep. 98-4148. Reston, VA.
926	http://pubs.usgs.gov/wri/wri984148/. Last Updated: 1 Sep 2005. Accessed 17 Nov 2010.
927	
928	Rutledge, A.T., 2007. Program User Guide for PART. U.S. Geological Survey.
929	http://water.usgs.gov/ogw/part/UserManualPART.pdf. Accessed 17 Nov 2010.
930	
931	Sacks, L.A., Swancar, A., and Lee, T.M., 1998. Estimating Ground-Water Exchange with Lakes
932	Using Water-Budget and Chemical Mass-Balance Approaches for Ten Lakes in Ridge Areas of
933	Polk and Highlands Counties, Florida. U.S. Geological Survey Water-Resour. Inv. Rep. 98-4133.
934	
935	Sanford, W.E., Nelms, D.L., Pope, J.P, and Selnick, D.L., 2012. Quantifying components of the
936	hydrological cycle in Virginia using chemical hydrograph separation and multiple regression
937	analysis. U.S. Geological Survey Sci. Inv. Rep. 2011-5198. 152p.
938	http://pubs.usgs.gov/sir/2011/5198/. Accessed 15 Jul 2014.
939	
940	Schindler, D.W., Newbury, R.W., Beaty, K.G., and Campbell, P., 1976. Natural Water and
941	Chemical Budgets for a Small Precambrian Lake Basin in Central Canada. J. Fish. Res. Board
942	Can. 33 (11): 2526–2543.

944

945	temperature of the free water surface. Soviet Hydrology, Selected Papers Issue 6.
946	
947	Singer, S.N., 1974. A Hydrogeological Study along the North Shore of Lake Ontario in the
948	Bowmanville-Newcastle Area. Ontario Ministry of the Environment, Water Resources Report
949	5d. Toronto, ON.
950	
951	Sloto, R.A., and Crouse, M.Y., 1996. HYSEP: A computer program for streamflow hydrograph
952	separation and analysis. U.S. Geological Survey Water-Resour. Inv. Rep. 96-4040, 46p.
953	
954	Stets, E.G., Winter, T.C, Rosenberry, D.O., and Striegl, R.G., 2010. Quantification of surface
955	water and groundwater flows to open- and closed-basin lakes in a headwaters watershed using a
956	descriptive oxygen stable isotope model. Water Resour. Res. 46 (3), W03515.
957	doi:10.1029/2009WR007793.
958	
959	SYKE, 2004. CORINE LAND COVER 2000, Pixel size 25m x 25m, clc_fi25m.tif. GIS digital
960	mapping data.
961	
962	SYKE, 2009. Aquifers [pohjavesialueet], PvesiAlue.shp and PvesiRaja.shp. GIS digital mapping
963	data.
	44

Shuliakovski, L.G., 1969. Formula for computing evaporation with allowance for the

964	9	6	4
-----	---	---	---

967	Taylor, J.R., 1997. An introduction to error analysis: the study of uncertainties in physical
968	measurements, second ed. University Science Books, Sausalito, CA.
969	
970	Therrien, R., McLaren, R.G., Sudicky, E.A. Panday, S., 2010. HydroGeoSphere. A three-
971	dimensional numerical model describing fully-integrated subsurface and surface flow and solute
972	transport. Draft User Manual. Groundwater Simul. Group, Univ. of Waterloo, Waterloo, ON.
973	http://www.ggl.ulaval.ca/fileadmin/ggl/documents/rtherrien/hydrogeosphere.pdf.
974	
975	Thodal, C.E., 1997. Hydrogeology of Lake Tahoe Basin, California and Nevada, and Results of a
976	Ground-Water Quality Monitoring Network, Water Years 1990-92. U.S. Geological Survey
977	Water-Resour. Inv. Rep. 97-4072.
978	
979	Trask, J.C., 2007. Resolving Hydrologic Water Balances through Novel Error Analysis, with
980	Focus on Inter-annual and long-term Variability in the Tahoe Basin. Ph.D. Dissertation,
981	University of California, Davis, CA, 378p.
982	
983	Tyler, F., 1977. A Laboratory Manual of Physics, fifth ed. Edward Arnold, London, U.K. 264p.
984	
	45

SYKE, 2010. Watershed areas [valuma-alueet], Jako3.shp. GIS digital mapping data.

985	Ventelä, AM., Arvola, L., Helminen, H., and Sarvala, J., 2005, Sense of place – meaning of a
000	$f_{\rm rescale}$ below for Einsich montal environment. Masitalous $4C(4)$, 27, 21. [In Einsich]
986	lake for Finnish mental environment. Vesitalous 46(4), 27–31. [In Finnish].
987	
988	Ventelä, AM., Tarvainen, M., Helminen, H., and Sarvala, J., 2007. Long-term management of
989	Pyhäjärvi (southwest Finland): eutrophication, restoration - recovery? Lake Reserv. Manage. 23,
990	428–438. doi:10.1080/07438140709354028.
991	
992	Walker, J. F., and Krabbenhoft, D.P., 1998. Groundwater and surface water interactions in
993	riparian and lake-dominated systems, in: Kendall, C., and McDonnell, J. J. (Eds.), Isotope
994	Tracers in Catchment Hydrology. Elsevier, New York, pp. 467–488.
995	
996	Wiebe, A.J., 2012. Quantifying the groundwater component within the water balance of a large
997	lake in a glaciated watershed: Lake Pyhäjärvi, SW Finland. M.Sc. Thesis, University of
998	Waterloo, Waterloo, ON, 143p. http://hdl.handle.net/10012/6490. Accessed 3 Aug 2013.
999	
1000	Winter, T.C., 1981. Uncertainties in estimating the water balance of lakes. Water Resour. Bull.
1001	17(1), 82–115.
1002	
1003	Winter, T.C., 1999. Relation of streams, lakes, and wetlands to groundwater flow systems.
1004	Hydrogeol. J. 7 (1), 28–45.
1005	

- 1006 Winter, T.C., and Rosenberry, D.O., 2009. Evaluation of methods and uncertainties in the water
- 1007 budget, in: Winter, T.C., and Likens, G.E. (Eds.), Mirror Lake: Interactions among air, land, and
- 1008 water. Berkeley, University of California Press, pp.205–224.

- 1010 Zacharias, I., Dimitriou, E., and Koussouris, T., 2003. Estimating groundwater discharge into a
- 1011 lake through underwater springs by using GIS technologies. Environ. Geol. 44, 843–851.
- 1012 doi:10.1016/j.jenvman.2003.09.017.
- 1013
- 1014 Zhang, Q., 2011. Development and application of an integrated hydrological model for lake
- 1015 watersheds. Procedia Environ. Sci. 10, 1630–1636. doi:10.1016/j.proenv.2011.09.257.

Table 1: Summary of equations and data used to calculate relative (rel) and absolute (abs) uncertainties for the Lake Pyhäjärvi water balance. Relative uncertainty = absolute uncertainty divided by the associated quantity.

Uncertainty Assigned in	Equations	Description of Uncertainty	Assigned	Source / Reference for
WB Equation			Uncertainty	Assigned Uncertainty
Lake Storage (h _s	.) Lake level via staff gauge OIVA/HERTTA (8 Sep 2010)	-		
$\delta h_{\rm S} = \pm 36 \rm mm$	-	Overall uncertainty on storage change [1]		
		Mananumant auroicieu alue lalte seiche		
	, 1#1,	Measurement precision plus lake selche effects	$0WL_{meas} = \pm 5.0$ mm	
	i"#\$96&'96/2 25 **	Lake seiche effects [2]	seiche = \pm 50 mm	Hyvärinen et al. (1973)
Lake Evaporatio	m (E) Class A Evaporation Pan (Jokioinen meteorological station, May - Sep) OIVA/	/HERTTA (5 Jun 2010)		
$\delta E = \pm 15\%$, +, -, /0 12312, - 455162,	Appropriateness of the assigned pan coefficient [3][4]	$\delta E_{rel} = [\delta E_{Jokio}]_{rel} = [\delta E_{Kuusisto}]_{rel} = 15\%$	Dingman (1994), Harbeck et al. (1954)
Pumping Withdr	"awals (W) Lohiluoma pumping station municipal records (J. Reko, pers. comm., 2010)			
$\delta W = \pm 11\%$	+, 7 81, 9 81, : ; +,	Lack of historical records on variable rates [5][6]	$\delta W_{daily} = \pm 500$ m ³ /d	
Lake Area (A)	Topographic and Bathymetric Elevations OIVA/HERTTA (10 Aug 2010), MML (2005	9b and 2009c)		
$\delta A = \pm 2.5 \text{ km}^2$; max ?@ ; AB ; AB ; iè D	Lake area dependence on lake stage [7]	Range of water level variation	
Precipitation (P)	Seven meteorological stations, areal interpolation method (OIVA/Hertta, 6 and 13 Oct	2010)		
$\delta P = \pm 5.2\%$ to $\pm 20\%$,+, 7 E5B 106+/2.6120,+, :	Instrument measurement errors plus spatial interpolation errors	$\delta P_{Gauge} = 5.0\%$	Winter (1981)
	,106 + /2,6120,+ , 12# 6, +, 1	Spatial interpolation over lake area [8]	0.2% to 15%	
Net River Discha	urge (R) Daily flow rates from weirs on Yläneenjoki, Pyhäjoki, and Eurajoki Rivers (OI	IVA/HERTTA, 23 Sep 2010), subwatershed	areas (SYKE, 2010; (DIVA/HERTTA,23Sep2010)
$\delta R = \pm 4.1\%$ to $\pm 89\%$, 7 ,0.6 : 7 G _{0,0} H : 7 L ₀ H : 7 J5,,H : *#0, ⁻ *#0, -	Overall river uncertainty [9]		
	"#0, $7 = \frac{1}{G_{n+1}}$; 7 ; +,B,+,: 7 ; "#0,+, : ; +, 96K +,	Extrapolating Yläneenjoki gauged flows to single channels [5][10]	$ \begin{split} & [\delta Reg]_{\rm nel}=\pm~9.0\% \\ & [\delta A_{\rm r,g}]_{\rm abs}=\pm1~{\rm km}^2 \\ & [\delta A_{\rm chan}]_{\rm abs}=\pm1{\rm km}^2 \end{split} $	Devito and Dillon (1993)
	$+_{n}H_{+}$, 7 $+_{n}+$, \vdots 7 ; $+_{B,+}$, $\ddot{7}$; $+_{6,+}$; \dot{z} $+,$	Each inflow river: final combined uncertainty [5][10][11]	$ \begin{bmatrix} \delta A_{r,g} \end{bmatrix}_{abs} = \pm 1 \text{ km}^2 \\ \begin{bmatrix} \delta A_{r,t} \end{bmatrix}_{abs} = \pm 1 \text{ km}^2 $	
	$J_{5,H,+}$, 7 $J_{5,+}$, \vdots $;$ +,	Outflow river: final combined uncertainty [5][9]		
	+, $LM = \frac{0}{1N} 7 + 1$:	Uncertainty associated with daily gauge readings [12][13]	$[\delta R_{r,i}]_{rel}=\pm~5.0\%$	Winter (1981), Herschy (1973)
Direct Runoff	Yläneenjoki R. daily flows (OIVA/HERTTA, 23 Sep 2010), surface geology (GTK, 2008)	(), land cover (SYKE, 2004), elevation (MM	L, 2009c), Barazzuoli	et al. (1989) coefficients.
<i>δDR</i> (Coeff. Map) = ±46% to ±50%	,+, 70 PQ.+,: 7,+,:	Estimating <i>DR</i> via Barazzuoli et al. (1989) runoff coefficients and watershed raster maps [14]	$\delta C_{DR} = 0.16$ $C_{DR} = 0.35$	Dhakal et al. (2012)
õDR (PART derived): ±13% to ±36%	$_{626,+,}$ L 7RST'U&VW $_{G,+,}$: $\%$ K +, Y $\frac{i}{i}$ Z Y $\frac{i}{i} \frac{PQ}{PQ}$ Z	Total error on PART derived <i>DR</i> [5][9][10]	$[\delta Reg]_{ m rel}=\pm~9.0\%$	Devito and Dillon (1993)
	RST'U&VV $\chi_{c_{u,+}}$, $[N] +, -[N] =$, RST'U&V $\chi_{K_{u,+}}$	Error on surface water fraction from PART (i.e., 1 – BFI) [9]	$[\delta BFI]_{ m rel} = \pm 8.2\%$	Sanford et al. (2012)

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Comments:

seiche effect. [3] Coefficient within range observed by Järvinen (1978). Jokioinen (subscript: Jokio) is the closest station to the lake. [4] Kuusisto (1975) assumed applying per unit area results from Yläneenjoki River to the *DR* areas. [11] A_{ri} is area of river *r*, total river watershed area. [12] R_{ri} indicates river *r*, daily reading *i*; *n* is the number of days in the year. [13] Uncertainty for a continuously monitored watershed. [14] G_{DR} is the runoff coefficient calculated for the *DR* regions. contour maps based on water levels observed between 1960 and 2010. [8] Differences between the isohyetal and areal average method estimates. [9] *YI* denotes Yläneenjoki River, Py denotes Pyhäjoki River, and Eu denotes Eurajoki River; f denotes final. The three terms in the square root are relative uncertainties. [10] 1] Combined uncertainty, where δW_{L_v} is the uncertainty of the water level at start of water year beginning in year y. [2] Uncertainty assigned to be half of the constant value of 8 mm per month for October to April based on a Dalton-type formula. Value assumed to have same uncertainty as the pan data. [5] A is the average lake area from OIVA/HERTTA. [6] Assigned uncertainty addresses variation in pumping and known minor withdrawals. [7] Areas obtained from A_{r,g} is area of river r, gauged watershed region only. A_{than} is the total area of the four single drainage channel regions. δReg is the regionalization error for $\begin{array}{c} 1018\\ 1020\\ 1022\\ 1022\\ 1022\\ 1023\\ 1025\\ 1025\\ 1026\\ 1027\\ 1027\\ 1026\end{array}$

1028 Table 2: Application of the runoff coefficient map method (Barazzuoli et al., 1989; Kennessey,1029 1930).

	Direct Runoff	Yläneenjoki River	Pyhäjoki River	
	Subwatersheds	Watershed	Watershed	
Permeability	Fraction of Subwatershed Area (Weight) ^a			
Bedrock	0.16 (0.30)	0.31 (0.30)	0.07 (0.30)	
Till	0.52 (0.25)	0.17 (0.25)	0.23 (0.25)	
Moraine ridges and hummocks	0.01 (0.20)	0.00 (0.20)	0.01 (0.20)	
Eskers and other glaciofluvial deposits	0.02 (0.05)	0.01 (0.05)	0.22 (0.05)	
Sand and gravel deposits	0.13 (0.10)	0.06 (0.10)	0.27 (0.10)	
Silt	0.06 (0.20)	0.05 (0.20)	0.06 (0.20)	
Clay	0.03 (0.25)	0.28 (0.25)	0.06 (0.25)	
Peat and organic matter	0.07 (0.20)	0.12 (0.20)	0.07 (0.20)	
Slope Angle		•		
0-3.5%	0.71 (0.03)	0.67 (0.03)	0.79 (0.03)	
3.5 - 10%	0.27 (0.05)	0.28 (0.05)	0.19 (0.05)	
10 - 35%	0.02 (0.20)	0.05 (0.20)	0.02 (0.20)	
> 35%	0.00 (N/A)	0.00 (0.30)	0.00 (0.30)	
Vegetative Cover		•		
Discontinuous urban fabric	0.08 (0.27)	0.01 (0.27)	0.00 (0.27)	
Industrial or commercial	0.01 (0.30)	0.00 (0.3)	0.00 (0.3)	
Sport and Leisure facilities	0.00 (0.25)	0.00 (0.25)	0.00 (0.25)	
Agriculture	0.13 (0.15)	0.34 (0.15)	0.27 (0.15)	
Forest	0.76 (0.05)	0.59 (0.15)	0.62 (0.05)	
Transitional woodland	0.02 (0.15)	0.06 (0.15)	0.10 (0.15)	
Mineral extraction	0.00 (N/A)	0.00 (N/A)	0.01 (0.3)	
Component ^b	Coefficients based on an Index of aridity > 0.40			
Permeability	0.23 0.25 0.16			
Slope angle	0.04	0.04	0.04	
Vegetation cover	0.09	0.09	0.09	
Sum (Fraction of annual <i>P</i> that is <i>DR</i>)	0.35	0.38	0.29	

1030 Notes:

a The weights (runoff coefficients related to the geology/slope/vegetation types) are based on the coefficients listed
 by Barazzuoli et al. (1989), who provide values for four categories for each component.

1033 ^b Component = physiographic theme

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Diverse Develop	Water Balance Component						
Direct Runon	(mm per year)						
Method Used for Water Balance	h _S Lake Level (Storage)	E Evaporation	W Pumping Withdrawals	P Precipitation	R Net River Flows	DR Direct Runoff	G Net groundwater Flow
			Component Aver	age			
Coefficient Map Method	3.2	450.1	11.1	607.7	-249.4	130.3	-24.3
PART Method	3.2	450.1	11.1	607.7	-249.4	63.1	43.0
Standard Deviation							
Coefficient Map Method	171.1	45.4	0.0	91.5	211.8	19.6	116.8
PART Method	171.1	45.4	0.0	91.5	211.8	26.9	118.0
Absolute Uncertainty							
Coefficient Map Method	36.1	67.5	1.2	61.2	21.8	61.2	119.4
PART Method	36.1	67.5	1.2	61.2	21.8	12.5	102.9

1037 Table 3: Summary of water balance results for water years 1971 to 2008 for two different1038 methods of calculating the direct runoff component.



Figure 1: The Lake Pyhäjärvi watershed and its two gauged, four single channel (i.e., ungauged), and six direct runoff (i.e., non-channelized) subwatersheds (imagery from MML, 2009a; ESRI, 2009; SYKE, 2010; and OIVA/HERTTA, 4 Jan 2011).



Figure 2: Surficial geology in the vicinity of Lake Pyhäjärvi, showing similarity between the Yläneenjoki River watershed and the non-channelized subwatersheds of the lake (after Wiebe, 2012; imagery from GTK, 2008; and SYKE, 2010).



Figure 3: Locations and names of coarse grained overburden aquifers in the vicinity of the Lake Pyhäjärvi watershed (after Wiebe, 2012; imagery from GTK, 2008; MML, 2009a; SYKE, 2009; and SYKE, 2010).





Figure 4: Water balance employing the coefficient map direct runoff estimate for water years 1971 to 2008. Components contributing a net gain to the lake are shown as positive; those exhibiting a net loss from the lake are shown as negative. Storage change is plotted above or below the zero line to indicate net gains or losses in storage. The error bars depict the annual absolute uncertainty values ($|\delta G|$) for the net groundwater flow.





Figure 5: Water balance employing the PART derived direct runoff estimate for water years 1971 to 2008. Components contributing a net gain to the lake are shown as positive; those exhibiting a net loss are shown as negative. Storage change is plotted above or below the zero line to indicate net gains or losses in storage. The error bars depict the annual absolute uncertainty values ($|\delta G|$) for the net groundwater flow.



Figure 6: A comparison of the net groundwater component magnitudes (|G|, expressed as positive values) to the absolute uncertainty values ($|\delta G|$) obtained for the PART derived water balances for water years 1971 to 2008. The actual values of *G* are shown for reference.





Figure 7: Baseflow index (BFI) values and corresponding direct runoff (*DR*) estimates calculated using PART for the Yläneenjoki and Pyhäjoki Rivers for water years 1971 to 2008. The *DR* value for Lake Pyhäjärvi from the coefficient map runoff method for the direct runoff subwatersheds is also shown for comparison.